

Original Research

Species Diversity of Carabid Beetles and Ants of Two Reclaimed Tailing Ponds

Mateusz Okrutniak*, Irena M. Grześ, Zbigniew Bonczar

Department of Environmental Zoology, Institute of Animal Science, University of Agriculture in Krakow, Kraków, Poland

Received: 6 September 2017

Accepted: 28 November 2017

Abstract

Metal pollution can drastically decrease the diversity of species of many groups of terrestrial invertebrates. Well-performed reclamation may prevent the loss of species diversity. In this study, we used multivariable statistical methods to describe the species diversity of carabids and ants originating from two zinc-lead mine tailing ponds under different degrees of reclamation practices. Although the species diversity of carabids and ants in both study sites and in both seasons was relatively high, the investigated communities were dominated by one numerous species. Discriminant analysis indicated that the difference between study sites is due to the high number of carabid species found in the comprehensively reclaimed site and the high abundance of several species of ants found in the partially reclaimed site. The most striking difference in species composition between the study sites was the contribution of forest species. The comprehensively reclaimed site maintains more species of forest carabids but fewer forest species of ants in comparison to the partially reclaimed one.

Keywords: metal pollution, carabids, ants, community

Introduction

Southern Poland is rich in zinc and lead ore deposits, and a long tradition of mining activities in this region dating back to the 14th century has led to the accumulation of large amounts of different types of waste [1]. Despite the fact that the technology for mining ores has been much improved, waste still makes up more than 90% of the output, causing pollution and land degradation [2]. The by-products of contemporary ore extraction are post-flotation wastes deposited in tailing ponds that mainly contain different fractions of dolomite, organic compounds, and trace metals, mainly

zinc, lead, cadmium, and copper. Because a large part of this waste is represented by a fraction of small dust, the hillsides and the dried-out top area of the tailing ponds are exposed to eolian erosion that may lead to further trace metal contamination of neighbouring areas [3].

Well-performed reclamation may prevent this process. In general, the remediation technologies of metal-contaminated soils consist of two alternatives. Contaminants can be removed from the substratum by, e.g., electrokinetic remediation, soil washing, or phytoextraction [4-6]. Alternatively, pollutants can be immobilized by, e.g., chemical stabilization and phytostabilization [7-8]. Due to the costs, time, and logistical constraints in the case of post-flotation wastes, the first alternative is rarely applied. The most common technique is phytostabilization, which is relatively

*e-mail: m.okrutniak@urk.edu.pl

inexpensive and environmentally sustainable and gives a long-term and visually attractive effect. Nevertheless, metalliferous wastes are very unfavourable environments for plants, and their introduction must be performed followed by additional procedures. First, the surface of the wastes must be stabilized and shaped (technical phase) and then they must be covered by a humus layer, which constitutes a fertile substrate for plants (biological phase) [3].

Carabid beetles and ants are considered pioneers in reclaimed areas. As Carson and Root [9] proved, the herbivorous activities of carabids can be crucial for plant diversity and dominance during the early phases of succession. In turn, both carabid beetles and ants as predators control the communities of many invertebrates and can influence decomposition and nutrient-cycling processes [10-11]. Additionally, the activity of ants improves the physical and chemical properties of soil [12-14].

The aim of the present study was to describe species diversity of carabids and ants inhabiting two separate, reclaimed zinc-lead tailing ponds named Trzebieńka and Bolesław. Both sites are extremely polluted by trace metals. Both tailing ponds are built of deposits of different ages. As a consequence, the plant cover represents variable stages of succession, from open areas to old forest stands. We endeavoured to collect samples of carabids and ants reflecting this heterogeneity; therefore, carabids and ants were sampled from various parts of the tailing ponds. Species diversity was described using multivariable statistical methods.

Materials and Methods

Study Sites

The present study was conducted at two tailing ponds: Trzebieńka and Bolesław. The process of storing the waste in both tailing ponds was similar. The upper parts of the tailing ponds represent the youngest deposits of waste material, while the bases represent the oldest deposits. The study sites differ in the effects of the reclamation processes. The hillsides of Trzebieńka have permanent plant cover representing different stages of succession, from early to advanced. Thus, the upper and lower parts of the pond are covered with planted herbaceous plants, shrubs, and trees, respectively. In turn, the hillsides of Bolesław tailing pond are at an early stage of succession and are dominated by herbaceous plants originating mainly from natural succession.

Trzebieńka tailing pond is located in Trzebinia (50°09'N, 19°25'E). It was established at the end of the 1960s and was in use until 2009. Trzebieńka has a trapezoidal shape with a 64 ha base and a height of about 40 m [15]. The basic properties of the post-flotation wastes of Trzebieńka are listed in Table 1. Both technical and biological reclamation work was simultaneously

applied to restore the hillsides. The technical reclamation was based on reinforcing the hillsides with a layer of crushed dolomite. The biological phase consisted of covering the waste with a 30-cm-deep humus layer, sowing grasses, and planting trees and shrubs that were selected for their high metal tolerance, drought tolerance, and low nutritional requirements. Additionally, the phytoremediation process was supported by a sprinkler system that watered the area during periods of drought [15].

The Bolesław tailing pond is located in Bolesław (50°17'N, 19°29'E). It was established in the 1950s and is still in use. The tailing pond has a trapezoidal shape and covers an area of 110 ha, rising above the surrounding terrain to a height of 42 m [17]. The basic properties of the post-flotation wastes of Bolesław are listed in Table 1. The technical reclamation conducted on Bolesław was based on reinforcing all hillsides with a layer of loamy deposit. The biological reclamation consisted of covering the waste with a 20-cm-deep layer of humus with municipal sewage sludge and sowing grasses. Unlike Trzebieńka, the plantings of trees and shrubs were not conducted simultaneously on all surfaces of the hillsides, but were introduced only to selected areas [18]. The small number of introduced plants and the lack of supporting the phytoremediation process by watering has led to the absence of permanent plant cover. In some small parts of the hillsides, planted species have disappeared and erosion has removed the soil over the waste, uncovering the substratum.

Sampling

The sampling of carabid beetles and ants was carried out on the hillsides of the tailing ponds from June to October in 2010 and 2011 (5 months in total). In both study sites, 60 pitfall traps filled with alcohol were installed. For each hillside exposition, i.e., north, south, and east, 20 pitfall traps were set up in two transects. One was placed on the upper part of the tailing ponds representing the youngest deposits (10 years), while the second transect was placed on the lower part representing the oldest deposits (40-50 years). The distance between the pitfall traps was 3 m, while the distance between the transects was about 100 m. The traps were exposed for the first three days of each month (June to October).

Table 1. Basic properties of post-flotation waste of the study areas, according to Trafas and Eckes [16].

		Trzebieńka	Bolesław
pH	H ₂ O	8.5	7.85
	KCl	8.4	7.83
Contents of components soluble in 60% HClO ₄ (mg/kg)	Zn	13,453	11,358
	Pb	2,936	2,271
	Cd	91	64.4

Table 2. Complete list of carabid beetles collected at the study sites.

Species	Trzebieńka		Bolesław	
	Total abundance	Relative abundance	Total abundance	Relative abundance
1. <i>Leistus (Pogonophorus) rufomarginatus</i> (Duftschmid, 1812).	0	0.00	2	0.50
2. <i>Leistus (Leistus) ferrugineus</i> (Linnaeus, 1758).	2	0.28	3	0.75
3. <i>Notiophilus aquaticus</i> (Linnaeus, 1758)	2	0.28	0	0.00
4. <i>Cicindela (Cicindela) campestris campestris</i> (Linnaeus, 1758)	0	0.00	2	0.50
5. <i>Cicindela (Cicindela) hybrida hybrida</i> (Linnaeus, 1758)	0	0.00	3	0.75
6. <i>Loricera pilicornis pilicornis</i> (Fabricius, 1775)	1	0.14	0	0.00
7. <i>Carabus (Megodontus) violaceus violaceus</i> (Linnaeus, 1758)	7	0.97	0	0.00
8. <i>Epaphius secalis secalis</i> (Paykull, 1790)	1	0.14	0	0.00
9. <i>Asaphidion pallipes</i> (Duftschmid, 1812)	5	0.69	2	0.50
10. <i>Metallina (Metallina) lampros</i> (Herbst, 1784)	27	3.72	14	3.52
11. <i>Metallina (Metallina) properans</i> (Stephens, 1828)	7	0.97	7	1.76
12. <i>Metallina (Chlorodium) pygmaeum</i> (Fabricius, 1792)	2	0.28	14	3.52
13. <i>Bembidion quadrimaculatum</i> (Linnaeus, 1761)	0	0.00	1	0.25
14. <i>Anisodactylus (Anisodactylus) binotatus</i> (Fabricius, 1787)	0	0.00	3	0.75
15. <i>Anisodactylus (Pseudanisodactylus) signatus</i> (Panzer, 1796)	1	0.14	0	0.00
16. <i>Pseudoophonus (Pseudoophonus) rufipes</i> (De Geer, 1774)	39	5.38	4	1.01
17. <i>Harpalus (Harpalus) affinis</i> (Schränk, 1781)	1	0.14	0	0.00
18. <i>Harpalus (Harpalus) atratus</i> (Latreille, 1804)	2	0.28	3	0.75
19. <i>Harpalus (Harpalus) distinguendus distinguendus</i> (Duftschmid, 1812)	3	0.41	0	0.00
20. <i>Harpalus (Harpalus) luteicornis</i> (Duftschmid, 1812)	1	0.14	0	0.00
21. <i>Harpalus (Harpalus) quadripunctatus</i> (Dejean, 1829)	4	0.55	3	0.75
22. <i>Harpalus (Harpalus) rubripes</i> (Duftschmid, 1812)	32	4.41	4	1.01
23. <i>Harpalus (Harpalus) serripes serripes</i> (Quensel in Schönherr, 1806)	3	0.41	1	0.25
24. <i>Harpalus (Harpalus) smaragdinus</i> (Duftschmid, 1812)	13	1.79	0	0.00
25. <i>Harpalus (Harpalus) tardus</i> (Panzer, 1796)	2	0.28	0	0.00
26. <i>Bradycellus (Bradycellus) caucasicus</i> (Chaudoir, 1846)	0	0.00	1	0.25
27. <i>Poecilus (Poecilus) cupreus cupreus</i> (Linnaeus, 1758)	4	0.55	0	0.00
28. <i>Poecilus (Makropoecilus) sericeus</i> (Fischer von Waldheim, 1824)	1	0.14	7	1.76
29. <i>Pterostichus (Platysma) niger</i> (Schaller, 1783)	16	2.21	0	0.00
30. <i>Pterostichus (Pseudomasesus) nigrita</i> (Paykul, 1790)	1	0.14	0	0.00
31. <i>Calathus (Neocalathus) ambiguus</i> (Paykull, 1790)	415	57.24	236	59.30
32. <i>Calathus (Neocalathus) erratus</i> (C, R, Sahlberg, 1827)	53	7.31	63	15.83
33. <i>Calathus (Neocalathus) melanocephalus</i> (Linnaeus, 1758)	1	0.14	0	0.00
34. <i>Anchomenus (Anchomenus) dorsalis</i> (Pontoppidan, 1763)	0	0.00	1	0.25
35. <i>Amara (Amara) eurynota</i> (Panzer, 1796)	5	0.69	1	0.25
36. <i>Amara (Amara) littorea</i> (C, G, Thomson, 1857)	17	2.34	8	2.01
37. <i>Amara (Amara) spreta</i> (Dejean, 1831)	15	2.07	2	0.50
38. <i>Amara (Celia) bifrons</i> (Gyllenhal, 1810)	20	2.76	0	0.00

Table 2. Continued.

39. <i>Amara (Xenocelia) cursitans</i> (C, Zimmermann, 1832)	7	0.97	1	0.25
40. <i>Amara (Percosia) equestris equestris</i> (Duftschmid, 1812)	6	0.83	2	0.50
41. <i>Amara (Curtonotus) aulica</i> (Panzer, 1796)	1	0.14	0	0.00
42. <i>Badister (Badister) bullatus</i> (Schrank, 1798)	2	0.28	0	0.00
43. <i>Panagaeus (Panagaeus) cruxmajor</i> (Linnaeus, 1758)	2	0.28	0	0.00
44. <i>Cymindis (Cymindis) humeralis</i> (Fourcroy, 1785)	0	0.00	8	2.01
45. <i>Microlestes minutulus</i> (Goeze, 1777)	4	0.55	2	0.50
In total	725	100	398	100

The traps were emptied each month during both sampling periods (2010 and 2011). The pooled sample of 10 traps consisting of one transect was used as a replicate in further analysis. The identification of carabids and ant workers was performed according to Hůrka [19] and Czechowski et al. [20], respectively. The identification of both groups was performed based on morphological traits, namely antennae location, antennae pubescence, shape of pronotum and elytra, body size in carabids, and number of waist segments, femur, body color, head shape in ants.

Statistical Analysis

In order to describe the fauna of carabid beetles and ants in the study sites, discriminant analysis was performed. Study site (Trzebieńka, Bolesław) was applied as a grouping variable, while species were used to design the discriminant functions. To simplify the discriminant function model, the forward selection method was applied. This means that variables (species) were added to the model one at a time. The final model contained only the significant variables of F, having a value of over 4.0. Carabid beetles and ants as well as sampling period (2010, 2011) were analysed separately. Each replicate (case) applied to the analysis consisted of the community of carabid beetles or ants collected from each transect in each month in 2010 or 2011. Thus 4 independent analyses

were performed: carabid beetles 2010, carabid beetles 2011, ants 2010, and ants 2011.

Additionally, in order to analyse the similarities in species composition between northern, southern, and eastern exposures we performed cluster analysis. To form the clusters, Ward's method with Euclidean distance metric was applied. The sampling period (2010, 2011) was analysed simultaneously in the case of carabid beetles as well as in the case of ants. All statistical analyses were performed with Statgraphics Centurion v. XVI (StatPoint Technologies, Inc., 2009).

Results and Discussion

Species Diversity of Carabid Beetles

In total, 1,123 specimens of carabid beetles belonging to 45 species were detected (Table 2) in Trzebieńka and Bolesław, which is about 9% of all species of carabids found in Poland [21]. In general, all identified species (except for *A. cursitans*) are common and widespread throughout Poland [21]. Both study sites were dominated by *C. ambiguus*, whose relative abundance amounted to 57% and 59% in Trzebieńka and Bolesław, respectively (Table 2). *C. ambiguus* prefers dry and warm unshaded habitats [19], which can explain its high abundance. Schwerk and Szyszko [22-23] also identified the high abundance of *Calathus* carabids in post-industrial areas. In terms of habitat preferences, the community of carabid beetles of both sites was established mainly by species of open areas, though Trzebieńka differed from Bolesław in terms of the higher number of forest carabids (Table 3). Most species identified in both study sites were of small or medium body size, which is in line with a similar study by Belskaya and Zolotarev [24] indicating that the contribution of small species of carabids is higher in disturbed than in undisturbed areas.

The discriminant analysis performed on carabid beetles collected in 2010 showed that 6 species – *A. bifrons*, *A. euronota*, *A. cursitans*, *P. niger*, *B. bullatus*, *P. cruxmajor* – can be used to design one function that significantly discriminates between Trzebieńka and

Table 3. Habitat preferences of carabid beetles and ants collected at the “Trzebieńka” and “Bolesław” study sites. Each number denotes the percentage share of species representing a certain habitat preference. The categories were defined according to Hůrka [19], Czechowski et al. [20], and Aleksandrowicz [21].

Category	Carabid beetles		Ants	
	Trzebieńka	Bolesław	Trzebieńka	Bolesław
Open areas	46	67	42	33
Forest or open areas	32	22	50	50
Forest	14	7	8	17
Riparian	8	4	-	-

Table 4. Results of discriminant analysis performed on carabid beetles collected in 2010. The listed species were selected by the analysis to construct a significant function discriminating between the “Trzebieńka” and “Bolesław” study sites. No discriminant function was designed from the data collected in 2011 due to the relatively low species F-values.

Species	F	Standardized coefficient
<i>A. bifrons</i>	7.43	0.75
<i>A. eurynota</i>	5.26	0.65
<i>A. cursistans</i>	4.98	0.59
<i>P. niger</i>	5.45	0.56
<i>B. bullatus</i>	5.33	0.51
<i>P. cruxmajor</i>	6.44	0.51

Bolesław ($P < 0.01$, Table 4). In 2010 these species were detected in Trzebieńka but not in Bolesław. Among these species, *A. bifrons* had the highest standardized coefficient (0.75, Table 4) being the most significant discriminator between Trzebieńka and Bolesław. Although a similar analysis was performed using data collected in 2011, no discriminant function was designed because of the relatively low species-F values. Cluster analysis showed that northern, southern, or eastern exposures did not

form distinct clusters (Fig. 1a), suggesting that exposure should not be considered a main factor shaping the variability of species diversity of carabids.

The higher number and abundance of carabids in Trzebieńka than in Bolesław (Table 2) can be explained by the fact that more trees and other plants (such as *Hippophae rhamnoides*, *Symphoricarpos albus*, and *Festuca ovina*) were introduced for reclamation purposes in Trzebieńka. This reclamation practice resulted in dense plant cover and hypothetically created more micro-climatically diverse habitats in Trzebieńka than in Bolesław within a relatively small area. In Trzebieńka, old forest stages of about 40-50 years are located at a distance of about 100 m from the areas covered by herbaceous plants (personal observation). According to Oliver et al., higher habitat heterogeneity provides more diverse resources and more differentiated niches that may meet the requirements of a much higher number of species than homogeneous habitats [25]. The heterogeneity-diversity relationship has been investigated in a number of studies, including beetles [26] and other insects [27-28]. However, not all trends are positive [29-30], which is hypothesized to be influenced by the differences in resource availability between the investigated communities [31]. A positive correlation between plant cover and number of carabid species is widely recognized [32-34]. Dense plant cover can retain humidity [35], supporting the establishment of mesohygrophilous or hygrophilous carabids such as *C. violaceus*, *A. signatus*, and *P. niger*, found in the present study only in Trzebieńka (Table 2).

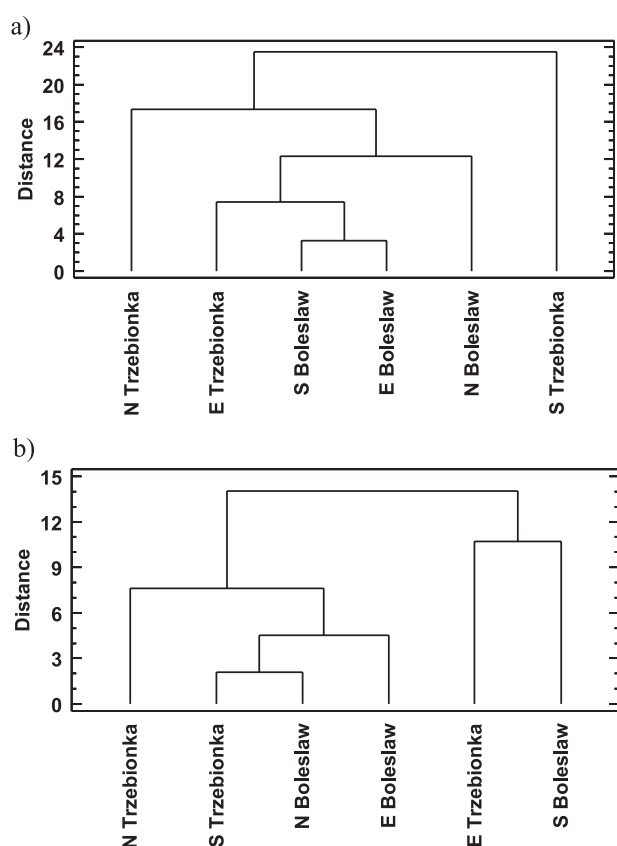


Fig. 1. The cluster analysis of the composition of species of carabid beetles a) and ants b) at the “Trzebieńka” and “Bolesław” study sites. N, S, E – denotes the exposure.

Species Diversity of Ants

In total, 6,108 specimens of ants representing 16 species were identified (Table 5), which is about 16% of all species of ants found in Poland [20]. In general, all species identified in the study sites are common and widespread in Poland [20]. The ant community was dominated by one species, namely *F. cinerea*, whose relative abundance accounted for more than 80% and 40% in Trzebieńka and Bolesław, respectively (Table 5). *F. cinerea* occupies thermophilic and xerophilic open habitats and forests, as well as sunny sandy sites with sparse vegetation [36].

As with the carabids, cluster analysis showed that exposure should not be considered a relevant factor that determines ant species diversity in investigated sites (Fig. 1b). In turn, the discriminant analysis applied to the ants collected in 2010 allowed us to construct one significant discriminating function based on two species: *M. rubida* and *F. cinerea* ($P = 0.01$, Table 6). When data collected in 2011 were analysed, one significant discriminant function was constructed that included 3 species, i.e., *M. ruginodis*, *L. niger*, and *M. rubida* ($P < 0.01$, Table 6). Based on an analysis of both seasons, *M. rubida* seems to be the most important discriminator between the study sites. In both seasons, this species was more abundant in Bolesław than in Trzebieńka. *M. rubida*

Table 5. Complete list of ants collected at the study sites.

Species		Trzebionka		Bolesław	
		Total abundance	Relative abundance	Total abundance	Relative abundance
1.	<i>Myrmica rubra</i> (Linnaeus, 1758)	17	0.53	0	0.00
2.	<i>Myrmica rugulosa</i> (Nylander, 1849)	14	0.44	17	0.59
3.	<i>Myrmica sabuleti</i> (Meinert, 1861)	1	0.03	0	0.00
4.	<i>Myrmica lonae</i> (Finzi, 1926)	2	0.06	0	0.00
5.	<i>Myrmica schencki</i> (Emery, 1895)	6	0.19	1	0.03
6.	<i>Manica rubida</i> (Latreille, 1802)	87	2.71	387	13.36
7.	<i>Stenamma debile</i> (Förster, 1850)	5	0.16	0	0.00
8.	<i>Tetramorium caespitum</i> (Linnaeus, 1758)	30	0.93	143	4.94
9.	<i>Formica (Formica) rufa</i> (Linnaeus, 1761)	0	0.00	170	5.87
10.	<i>Formica (Formica) truncorum</i> (Fabricius, 1804)	0	0.00	19	0.66
11.	<i>Formica (Serviformica) fusca</i> (Linnaeus, 1758)	10	0.31	1	0.03
12.	<i>Formica (Serviformica) cinerea</i> (Mayr, 1853)	2,670	83.15	1,366	47.15
13.	<i>Formica (Serviformica) cunicularia</i> (Latreille, 1798)	19	0.59	2	0.07
14.	<i>Formica (Raptiformica) sanguinea</i> (Latreille, 1798)	0	0.00	40	1.38
15.	<i>Lasius (Lasius) niger</i> (Linnaeus, 1758)	350	10.90	750	25.89
16.	<i>Lasius (Chthonolasius) distinguendus</i> (Emery, 1916)	0	0.00	1	0.03
In total		3,211	100	2,897	100

is montane species frequent in Central and south-eastern Europe occurring at 500-2,000 m a.s.l. [36]. Even though this species is considered typical for mountain grassy habitats, it was also recorded in the lowlands and also on coal mining spoil dumps [37]. It was concluded that *M. rubida* is an efficient colonizer of disturbed areas [37].

In terms of habitat preferences, the community of ants of both sites was established mainly by forest and open-area-species (e.g. *M. schencki*, *F. cinerea*, *L. niger*),

but more forest species, such as *F. rufa*, *F. truncorum*, were found in Bolesław than in Trzebionka (Table 3). The forest species detected in Bolesław may be foragers from the coniferous forest located close to the study area, as species such as *F. rufa* can forage over distances greater than 100 m [20].

Conclusions

To summarize, although the investigated sites are relatively small and highly polluted by trace metals, the species diversity of both carabids and ants in both study sites and both seasons is relatively high. However, the communities are established by common and widespread species in Poland and are highly dominated. The composition of species of carabid beetles and ants differed between the investigated sites. We suggest that these results reflect the reclamation practices that created variable microhabitats. This variability shapes the differences in the compositions of species, most likely due to the presence of more humid niches in Trzebionka, the comprehensively reclaimed tailing pond. The difference between the study sites was shown to be more important in shaping the species diversity of both carabids and ants than northern, southern, or eastern exposure.

Table 6. The results of the discriminant analysis performed on ants collected in 2010 and 2011. The listed species were selected by the analysis to construct a significant function discriminating between the “Trzebionka” and “Bolesław” study sites.

Species	F	Standardized coefficient
2010		
<i>M. rubida</i>	4.59	0.74
<i>F. cinerea</i>	4.01	-0.70
2011		
<i>M. ruginodis</i>	5.85	0.85
<i>L. niger</i>	4.50	-0.60
<i>M. rubida</i>	4.88	-0.88

Acknowledgements

This study was financed by the Ministry of Science and Education (MNiSW), DS 3247/ZZŚ. We thank Tomasz Skalski for his help in the identification of carabids. Author contributions: MO designed the study and performed the experiment, IG participated in the identification of ants, and all authors contributed to the writing of the manuscript.

References

- KICIŃSKA A., GRUSZECKA-KOSOWSKA A. Long-term changes of metal contents in two metallophyte species (Olkusz area of Zn-Pb ores, Poland). *Environ. Monit. Assess.*, **188**, 339, **2016**.
- KARCZEWSKA A. Soil protection and reclamation of degraded areas Wydawnictwo Uniwersytetu Przyrodniczego we Wrocławiu, Wrocław, Poland, **2012** [In Polish].
- TORDOFF G.M., BAKER A.J.M., WILLIS A.J. Current approaches to the revegetation and reclamation of metalliferous mine wastes. *Chemosphere*, **41**, 219, **2000**.
- ROSA M.A., EGIDO J.A., MÁRQUEZ M.C. Enhanced electrochemical removal of arsenic and heavy metals from mine tailings. *J. Taiwan Inst. Chem. E.*, **78**, 409, **2017**.
- KARLFELDT FEDJE K., YILLIN L., STRÖMVALL A.M. Remediation of metal polluted hotspot areas through enhanced soil washing-Evaluation of leaching methods. *J. Environ. Manage.*, **128**, 489, **2013**.
- JEELANI N., YANG W., XU L., QIAO Y., AN S., LENG X. Phytoremediation potential of *Acorus calamus* in soils co-contaminated with cadmium and polycyclic aromatic hydrocarbons. *Sci. Rep.-Uk.*, **7**, 8028, **2017**.
- KOMÁREK M., VANĚK A., ETTLER V. Chemical stabilization of metals and arsenic in contaminated soils using oxides-A review. *Environ. Pollut.*, **172**, 9, **2013**.
- GU H.H., ZHOU Z., GAO Y.Q., YUAN X.T., AI Y.J., ZHANG J.Y., ZUO W.Z., TAYLOR A.A., NAN S.Q., LI F.P. The influences of arbuscular mycorrhizal fungus on phytostabilization of lead/zinc tailings using four plant species. *International Journal of Phytoremediation*, **19** (8), 739, **2017**.
- CARSON W.P., ROOT R.B. Top-down effects of insect herbivores during early succession: influence on biomass and plant dominance. *Oecologia*, **121**, 260, **1999**.
- HUNTER M.D., ADL S., PRINGLE C.M., COLEMAN D.C. Relative effects of macro invertebrates and habitat on the chemistry of litter during decomposition. *Pedobiologia*, **47**, 101, **2003**.
- GOLLAN J.R., DE BRUYN L.L., REID N., SMITH D., WILKIE L. Can ants be used as ecological indicators of restoration progress in dynamic environments? A case study in a revegetated riparian zone. *Ecological Indicators*, **11**, 1517, **2011**.
- EHRLE A., ANDERSEN A.N., LEVICK S.R., SCHUMACHER J., TRUMBORE S.E., MICHALZIK B. Yellow-meadow ant (*Lasius flavus*) mound development determines soil properties and growth responses of different plant functional types. *Eur. J. Soil Biol.*, **81**, 83, **2017**.
- JÍLKOVÁ V., PECH P., MIHALJEVIČ M., FROUZ J. Effects of the ants *Formica sanguinea*, *Lasius niger*, and *Tetramorium cf. caespitum* on soil properties in an ore-washery sedimentation basin. *J. Soils Sediments*, **17**, 2127, **2017**.
- WANG C., WANG G., WU P., RAFIQUE R., ZI H., LI X., LUO Y. Effects of ant mounds on the plant and soil microbial community in an alpine meadow of Qinghai-Tibet plateau. *Land Degrad. Develop.*, **28**, 1538, **2017**.
- KŁOJZY-KARCZMARCZYK B., MAZUREK J., CHOBOT C., MAKOUDI S., ŻÓLTEK J., KUREK T. Technical documentation of the termination and reclamation of Z.G. "Trzebionka" S.A. tailing pond. Instytut Gospodarki Surowcami Mineralnymi i Energią PAN, Kraków, Poland, **2009** [In Polish].
- TRAFAS M., ECKES T. Soil-making aspects in the evaluation of artificial formations; focus on the wastes formed after the flotation of zinc and lead ores. *Geomatics and Environmental Engineering*, **1**, 97, **2007** [In Polish].
- BAUEREK A., BEBEK M., SRACEK O., SMIEJA-KROL B. Chemical composition of surface runoff from flotation wastes of Zn-Pb ore formation of the Mississippi Valley-type, Olkusz, Southern Poland. *J. Geochem. Explor.*, **132**, 54, **2013**.
- KUCZYŃSKA I. Technical report on the influence of the elevation of ZGH "Bolesław" S.A., Kraków, Poland, **2009** [In Polish].
- HŮRKA K. Carabidae of Czech and Slovak Republics; Kabourek, Zlín, Czech Republic, **1996**.
- CZECHOWSKI W., RADCHENKO A., CZECHOWSKA W., VEPSÄLÄINEN K. The ants of Poland with reference to the myrmecofauna of Europe; Museum and Institute of Zoology, Polish Academy of Sciences, Natura optima dux Foundation, Warszawa, Poland, **2012**.
- ALEKSANDROWICZ O.R. Carabids (Carabidae), in: The Fauna of Poland – characteristics and the list of species. T.1; Bogdanowicz W., Chudzińska E., Filipiuk I., Skibińska E. (Eds.), MiIZ PAN, Warszawa, Poland, **28**, **2004** [In Polish].
- SCHWERK A., SZYSZKO J. Succession of carabid fauna (Coleoptera: Carabidae) on post-industrial areas near Bełchatów (Central Poland). *Wiadomości Entomologiczne*, **25**, 71, **2006**.
- SCHWERK A. Changes in carabid beetle fauna (Coleoptera: Carabidae) along successional gradients in post-industrial areas in Central Poland. *Eur. J. Entomol.*, **111** (5) 677, **2014**.
- BELSKAYA E.A., ZOLOTAREV M.P. Changes in the size structure of carabid communities in forest ecosystems under technogenic transformation. *Russ. J. Ecol.*, **48** (2), 152-, **2017**.
- OLIVER T., ROY D.B., HILL J.K., BRERETON T., THOMAS C.D. Heterogeneous landscapes promote population stability. *Ecol. Lett.*, **13**, 473, **2010**.
- NEUMANN J.L., GRIFFITHS G.H., HOODLESS A., HOLLOWAY G.J. The compositional and configurational heterogeneity of matrix habitats shape woodland carabid communities in wooded-agricultural landscapes. *Landscape Ecol.*, **31**, 301, **2016**.
- MALLINGER R.E., GIBBS J., GRATTON C. Diverse landscapes have a higher abundance and species richness of spring wild bees by providing complementary floral resources over bees' foraging periods. *Landscape Ecol.*, **31**, 1523, **2016**.
- PAPANIKOLAOU A.D., KÜHN I., FRENZEL M., SCHWEIGER O. Landscape heterogeneity enhances stability of wild bee abundance under highly varying temperature, but not under highly varying precipitation. *Landscape Ecol.*, **32**, 581-593, **2017**.

29. BAR-MASSADA A., WOOD E.M. The richness-heterogeneity relationship differs between heterogeneity measures within and among habitats. *Ecography*, **37**, 528, **2014**.
30. GAZOL A., TAMME R., PRICE J.N., HIIESALU I., LAANISTO L., PÄRTEL M. A negative heterogeneity-diversity relationship found in experimental grassland communities. *Oecologia*, **173**, 545, **2013**.
31. YANG Z., LIU X., ZHOU M., AI D., WANG G., WANG Y., CHU C., LUNDHOLM J.T. The effect of environmental heterogeneity on species richness depends on community position along the environmental gradient. *Sci. Rep.-Uk.*, **5**, 15723, **2015**.
32. BUTTERFIELD J. Carabid community succession during the forestry cycle in conifer plantations. *Ecography*, **20**, 614, **1997**.
33. MAGURA T., TOTHMERSZ B. Testing edge effect on carabid assemblages in an oak-hornbeam forest. *Acta Zool. Hung.*, **43**, 303, **1997**.
34. MAGURA T., TOTHMERSZ B., MONAR T. Spatial distribution of carabids along grass-forest transects, *Acta Zool. Hung.*, **46**, 1, **2000**.
35. TRAFAS M. Changes in the properties of post-flotation wastes due to vegetation introduced during process of reclamation. *Appl. Geochem.*, **11**, 181, **1996**.
36. SEIFERT B. *Die Ameisen Mittel- und Nordeuropas*, Lutra Verlags und Vertriebsgesellschaft, Görlitz, Germany, **2007**.
37. HOLEC M., FROUZ J. Ant (Hymenoptera: Formicidae) communities in reclaimed and unreclaimed brown coal mining spoil dumps in the Czech Republic. *Pedobiologia*, **49**, 345, **2005**.